Characteristics of the NO-NO$_2$-O$_3$ system in different chemical regimes during the MIRAGE-Mex field campaign


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Abstract. The NO-NO$_2$ system was analyzed in different chemical regimes/air masses based on observations of reactive nitrogen species and peroxy radicals made during the intensive field campaign MIRAGE-Mex (4 to 29 March 2006). The air masses were categorized into 5 groups based on combinations of macroscopic observations, geographical location, meteorological parameters, models, and observations of trace gases: boundary layer (labeled as “BL”), biomass burning (“BB”), free troposphere (continental, “FTCO” and marine, “FTMA”), and Tula industrial complex (“TIC”). In general, NO$_2$/NO ratios in different air masses are near photo-stationary state. Analysis of this ratio can be useful for testing current understanding of tropospheric chemistry. The ozone production efficiency (OPE) for the 5 air mass categories ranged from 4.5 (TIC) to 8.5 (FTMA), consistent with photochemical aging of air masses exiting the Mexico City Metropolitan Area.

1 Introduction

During March of 2006, the Megacities Impact on Regional and Global Environment: Mexico (MIRAGE-Mex) field campaign took place over the Mexico City Metropolitan Area (MCMA) and the surrounding region including the Gulf of Mexico, as part of the MILAGRO Project. This field campaign was designed to examine the chemical and physical transformations of gases and aerosols in the polluted outflow from MCMA and to assess the current and future impacts of these exported pollutants on regional and global air quality, ecosystems, and climate. MCMA is a megacity of about 20 million people residing in an area of 1500 km$^2$, surrounded by mountains and at an elevation of 2.2 km above sea level (a.s.l.). The air quality in MCMA is affected by strong anthropogenic sources of NO$_x$ (NO+NO$_2$) and volatile organic compounds (VOCs) in conjunction with high solar irradiance facilitating photochemistry (Raga et al., 2001). In the 1990’s, hourly averaged ozone concentrations in MCMA exceeded the Mexican national standard of 110 ppbv for much of the year (Raga and Raga, 2000).

As essential ingredients for the formation of O$_3$, the reactive nitrogen species (NO$_x$) are emitted to the atmosphere mainly in the form of NO from road traffic (Soltic and Weilenmann, 2003). NO and NO$_2$ are interconverted rapidly through the following reactions:

$$\text{NO} + \text{O}_3 \rightarrow \text{NO}_2 + \text{O}_2 \quad (R1)$$
$$\text{NO}_2 + h\nu (\lambda < 420 \text{ nm}) (+\text{O}_2) \rightarrow \text{NO} + \text{O}_3 \quad (R2)$$
$$\text{NO} + \text{HO}_2 \rightarrow \text{NO}_2 + \text{OH} \quad (R3)$$
$$\text{NO} + \text{RO}_2 \rightarrow \text{NO}_2 + \text{RO} \quad (R4)$$
$$\text{NO} + \text{XO} \rightarrow \text{NO}_2 + \text{X} \quad (R5)$$

where RO$_2$ is any organic peroxy radical including CH$_3$O$_2$, RO is the corresponding alkoxy radical, and XO is any unknown radical that may convert NO to NO$_2$. The NO$_2$/NO ratio can be formulated as:

$$\frac{[\text{NO}_2]}{[\text{NO}]} = \frac{(k_1[\text{O}_3] + k_3[\text{HO}_2] + k_4[\text{RO}_2] + k_5[\text{XO}])}{J_2} \quad (1)$$

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where $k_1$, $k_3$, $k_4$, and $k_5$ are reaction rate constants for the reactions R1, R3, R4, and R5, respectively and $J_2$ is the photolysis frequency of NO$_2$. Previous studies investigating the NO-NO$_2$ cyclic system were conducted in urban and remote areas utilizing observations of O$_3$ (R1) and/or peroxy radicals (R3–R4) (Cantrell et al., 1997; Crawford et al., 1996 and references therein). They found that model-predicted peroxy radicals were often less than those required to explain the observed NO$_2$/NO ratio, resulting in model-predicted NO$_2$ levels that were somewhat lower than observations. For example, comparison between observations and predictions in several field campaigns (CITE-3, ABLE-3B, CITE-2, and TRACE-A) typically showed that [NO$_2$]$_{obs}$/[NO$_2$]$_{calc}$ ranges from 1.3 to 1.6 with a maximum of up to 3.4 in PEM-West A (Crawford et al., 1996). In contrast, Ridley et al. (1992) found good agreement between model peroxy radicals and those estimated from the NO$_2$/NO ratio. Meanwhile, the potential role of iodine chemistry on the NO$_2$ formation has been suggested based on field, model, and kinetic studies (Chameides and Davis, 1980; Davis et al., 1996; Knight and Crowley, 2001; Kanaya et al., 2002, 2007). Since both hydroperoxyl and organic peroxy radicals were measured during the MIRAGE-Mex campaign, the photostationary state (PSS) of the NO-NO$_2$ system can be assessed based on field observations in this study without model estimates.

To assess the impact of MCMA emissions on O$_3$ and its precursors on regional and hemispheric scales, it is important to evaluate the loss and transformation processes of reactive nitrogen species (NO$_x$ to NO$_y$, which is total reactive nitrogen oxides) in the course of transport of urban or industrial plumes. The emitted NO$_x$ can be oxidized in the atmosphere by OH, forming HNO$_3$ which is subject to removal from the air mass through dry and wet deposition. The average lifetime of NO$_x$ in urban, industrial (Nunnermacker et al., 2000) or continental outflow plumes (Takegawa et al., 2004) ranges from 0.25 to several days; whereas that of NO$_x$ in each environment is slightly shorter (less than 0.25 day in the former environment with longer lifetime in the latter environment e.g., 1–2 days).

In this study, we analyzed the NO-NO$_2$ cyclic system in different chemical regimes/air masses based on observations of reactive nitrogen species and peroxy radicals. The analysis of PSS allows us to assess the current understanding of tropospheric NO$_x$ chemistry including the potential for yet-unidentified chemical reactions. We also examined the NO$_x$/NO$_y$ ratios and ozone production efficiencies (OPEs) of the polluted outflow from MCMA, providing some insight on the photochemical aging processes in various chemical regimes.

## 2 Observational data

The intensive field campaign of MIRAGE-Mex was carried out from 4 March to 29 March 2006, involving the NSF/NCAR C-130 aircraft, ground-based measurements, and satellite observations. This campaign made numerous physico-chemical measurements such as reactive nitrogen species, oxidized sulfur species, oxygenated VOCs, aerosols, peroxy radicals, and so on. It also included the measurements of actinic flux and atmospheric photolysis frequencies such as $J$(O$_3$), $J$(NO$_2$), $J$(HNO$_2$), etc., with 5-s time resolution (Shetter et al., 2003). These measurements allow for the PSS analysis of the NO-NO$_2$ system and are also used for air mass classification. The full details of measurements taken from C-130 during the field campaign can be found elsewhere (http://mirage-mex.ucar.edu/Measurements/C130/index.shtml). Reactive nitrogen species, including NO, NO$_2$, NO$_y$, and O$_3$ were measured at 1 Hz with a chemiluminescence technique (Weinheimer et al., 1998). For NO, NO$_2$, and NO$_y$, the precision of these measurements is near 15 pptv and the overall estimated uncertainties are ±(15–7% of the mixing ratio) pptv for NO, ±(15–10% of the mixing ratio) pptv for NO$_2$, and ±(15–15% of mixing ratio) pptv for NO$_y$. As one of key measurements for the analysis of NO-NO$_2$ PSS (Eq. 1), peroxy radicals (HO$_2$+RO$_2$) were measured with the four-channel CIMS at 3-s time interval (Cantrell et al., 2003). The overall estimated uncertainty of the peroxy radical measurement is ±(2.5+17.5% of the mixing ratio) pptv. In this study, 1-min merged data sets were used, which were created to align chemical species measured with varying time resolution and are available on MILAGRO/INTEX-B/IMPEX archive (ftp://ftp-air.larc.nasa.gov).

Twelve C-130 missions were flown during the campaign, covering the altitude from the surface to about 7 km. Most of the flights sampled air over the MCMA basin and central Mexico, and several of the flights extended over the Gulf of Mexico to sample continental outflow from the MCMA. Figure 1 shows C-130 flight tracks during the field campaign. In the temporal coverage, data sets used in this study were measured predominantly during daytime and (in smaller number) near sunrise and sunset. For this analysis, data were limited to solar zenith angles (SZA) less than 85°. The data of flight 6 were excluded in the PSS analysis due to malfunction of the ozone instrument.

## 3 Results and discussion

### 3.1 Air mass category

In order to analyze the NO-NO$_2$ cycling in different chemical regimes, the air masses were categorized into 5 groups: boundary layer (BL), free troposphere (continental, FTCo and marine, FTMA), biomass burning (BB), and Tula Industrial Complex (TIC). The air mass characterization was determined based on combinations of macroscopic observations, geographical location, meteorological parameters (temperature and relative humidity, etc.), models (Weather Research...
Fig. 1. C-130 flight tracks during the MIRAGE-Mex field campaign (March 2006).

Table 1. Criteria for the determination of air mass category.

<table>
<thead>
<tr>
<th>Altitude Geographical Location</th>
<th>HCN (pptv)</th>
<th>C$_2$Cl$_4$ (pptv)</th>
<th>CO (ppbv)</th>
<th>NO$_x$ (ppbv)</th>
<th>SO$_2$ (ppbv)</th>
</tr>
</thead>
<tbody>
<tr>
<td>BL 0.9 (F4)–1.2 (F1)*</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>FTCO &gt;{0.9 (F4)–1.2 (F1)}</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>FTMA &gt;{0.9 (F4)–1.2 (F1)}</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>BB ≥189</td>
<td>≥0.3</td>
<td>≥81</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>TIC around Tula Refinery</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

* The number represents the range of minimum and maximum values for median values of BL heights of each flight (F1 to F12).

and Forecasting (WRF) with Chemistry, WRF-Chem, Tie et al., 2007 and references therein), and observations of trace gases (Table 1). For instance, the category for the BL air masses was determined based on meteorological parameters and WRF-Chem and those for FTCO and FTMA were based on geographical locations of airborne sampling. In addition, that for the BB was determined based on macroscopic observation (i.e., fire in nearby mountains) and chemical observations (i.e., significant increase in concentration levels compared to previous flight legs) such as hydrogen cyanide (HCN), perchlorethene (C$_2$Cl$_4$), and CO (Gregory et al., 1996; Li et al., 2000), while that for TIC was based on the concentration levels of CO, NO$_x$, and SO$_2$. A statistical summary of the concentrations observed in the different air mass categories is given in Table 2. The mean concentration of HCN for the BB (measured on 4, 22, and 23 March) is factors of 1.5 (TIC), 1.8 (BL), 2.4 (FTMA), and 2.9 (FTCO) higher than those for other air mass categories. The mean concentration of SO$_2$ for the TIC (measured on 10, 19, and 22 March) is higher than those for other air mass categories by factors ranging from 4.7 to 16. High levels of NO$_x$ and NO$_y$ within the megacity result from enhanced local emission sources such as combustion. Thus, NO$_x$ can be a measure of anthropogenic impact at the sampling position. In addition, the ratio of NO$_x$ to NO$_y$ can be valuable for understanding the extent of chemical processing and as an indicator for photochemical age (Carroll et al., 1992), in spite of several limitations including the assumptions of no physicochemical loss (only dilution of NO$_y$), no chemical transformation from NO$_y$ to NO$_x$ (e.g., thermal decomposition of PAN), constant emission sources, and constant background concentration. In this field study, the mean NO$_x$/NO$_y$ ratio varied from 0.15 (FTMA) to 0.38 (BL), depending on air mass category (Table 2). The higher mean ratios (0.21–0.38) for BL, BB, and TIC (Table 2) indicate less aged air than those (0.15–0.19) for FTCO and FTMA.
Fig. 2. Relation between NO$_x$/NO$_y$ and NO$_y$ for 5 air mass categories (BL a, BB b, FTCO c, FTMA d, and TIC e).

(Table 2). The slightly lower ratios for FTCO and FTMA indicated more photochemical aging.

The variation of NO$_x$/NO$_y$ ratios with concentration level of NO$_y$ for each air mass is shown in Fig. 2, which can provide information on the characteristics of different air masses. In general, some positive correlation might be expected because high NO$_y$ may indicate proximity to sources and high NO$_x$/NO$_y$ ratios indicate freshly emitted pollution before much photochemical processing has occurred. For the BL, the ratios significantly varied from 0.1 to 1 with high NO$_y$ concentration, indicating a range of photochemical ages. In contrast, the ratios for the BB were concentrated at 0.2, with only a few values above 0.4. In general, the air masses for FTCO and FTMA appear to have similarity in the ratio patterns, with relatively low ratios (aged air) for NO$_y$ levels larger than a few ppbv, indicating that the polluted plumes have aged substantially. However, the ratios for FTCO at low NO$_y$ levels (less than a few ppbv) showed larger variation than the ratios for FTMA and those at high NO$_y$ levels (greater than a few ppbv) were also somewhat lower. Note that the high ratios at low absolute NO$_y$ were not affected by the low NO$_y$ levels which are significantly higher than NO$_y$ detection limit. The ratios for the TIC correlate fairly well with NO$_y$, as might be expected from dilution and aging of the plume from this fairly localized source.

The NO$_x$/NO$_y$ ratios during MIRAGE-Mex were significantly higher than those (0.02–0.2) in continental outflow from East Asia during TRACE-P (Koike et al., 2003). However, the higher ratio (0.38) for BL was comparable to those in the near 2 day aged Asian outflow plume measured during the PEACE-A campaign (Takegawa et al., 2004). Our ratios (0.15–0.19) for the FTCO and FTMA were similar to
Table 2. A statistical summary of reactive nitrogen compounds, major air pollutants, and peroxy radicals with air mass categories measured during the MIRAGE field campaign.

<table>
<thead>
<tr>
<th>Air mass category</th>
<th>NO (ppbv)</th>
<th>NO2(ppbv)</th>
<th>NOx (ppbv)</th>
<th>NOy (ppbv)</th>
<th>CO (ppbv)</th>
<th>O3 (ppbv)</th>
<th>SO2 (ppbv)</th>
<th>HO2 (pptv)</th>
<th>RO2 (ppbv)</th>
<th>HCN (pptv)</th>
<th>ClCl4 (pptv)</th>
<th>NOx/NOy</th>
<th>NOx/NOy^a</th>
</tr>
</thead>
<tbody>
<tr>
<td>BL</td>
<td>0.7±1.1^b</td>
<td>3.0±4.2</td>
<td>3.7±5.1</td>
<td>7.8±7.4</td>
<td>249±110</td>
<td>79±25</td>
<td>4.7±17.9</td>
<td>77±49</td>
<td>54±48</td>
<td>529±225</td>
<td>8.6±9.5</td>
<td>0.38±0.22</td>
<td>4.4±4.2</td>
</tr>
<tr>
<td>0.2^b</td>
<td>1</td>
<td>1.3</td>
<td>4.9</td>
<td>223</td>
<td>76</td>
<td>1.9</td>
<td>55</td>
<td>35</td>
<td>475</td>
<td>3</td>
<td>0.09±0.17</td>
<td>1.5±1.7</td>
<td></td>
</tr>
<tr>
<td>0.01–1.86^b,6</td>
<td>0.07–27</td>
<td>0.09–32</td>
<td>0.9–37</td>
<td>111–688</td>
<td>32–153</td>
<td>0.2–207</td>
<td>12–246</td>
<td>0.8–241</td>
<td>96–1291</td>
<td>0.5–31</td>
<td>0.09–1.47</td>
<td>1.5–17</td>
<td></td>
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<tr>
<td>387^d</td>
<td>387</td>
<td>387</td>
<td>375</td>
<td>170</td>
<td>396</td>
<td>278</td>
<td>267</td>
<td>255</td>
<td>291</td>
<td>126</td>
<td>375</td>
<td>387</td>
<td></td>
</tr>
<tr>
<td>BB</td>
<td>0.5±0.8</td>
<td>2.5±4.6</td>
<td>3.1±5.3</td>
<td>10.2±11.5</td>
<td>278±184</td>
<td>74±30</td>
<td>2.7±2.1</td>
<td>58±26</td>
<td>159±71</td>
<td>965±761</td>
<td>4.2±7.5</td>
<td>0.21±0.14</td>
<td>4.1±1.5</td>
</tr>
<tr>
<td>0.2</td>
<td>0.7</td>
<td>0.9</td>
<td>6.3</td>
<td>225</td>
<td>72</td>
<td>2</td>
<td>59</td>
<td>158</td>
<td>905</td>
<td>0.9</td>
<td>0.15</td>
<td>3.7</td>
<td></td>
</tr>
<tr>
<td>0.02–4.0</td>
<td>0.06–23</td>
<td>0.08–26</td>
<td>0.7–59</td>
<td>81–937</td>
<td>29–151</td>
<td>0.2–97</td>
<td>12–106</td>
<td>58–362</td>
<td>189–5952</td>
<td>0.3–32</td>
<td>0.07–0.71</td>
<td>1.7–10</td>
<td></td>
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<tr>
<td>88</td>
<td>88</td>
<td>88</td>
<td>86</td>
<td>58</td>
<td>90</td>
<td>60</td>
<td>24</td>
<td>30</td>
<td>65</td>
<td>30</td>
<td>86</td>
<td>88</td>
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<tr>
<td>FTCO</td>
<td>0.05±0.9</td>
<td>0.2±0.4</td>
<td>0.3±0.5</td>
<td>1.7±2.6</td>
<td>128±41</td>
<td>51±14</td>
<td>2.0±6.8</td>
<td>45±33</td>
<td>38±153</td>
<td>1.3±3.9</td>
<td>0.19±0.12</td>
<td>4.2±4.3</td>
<td></td>
</tr>
<tr>
<td>0.05</td>
<td>0.1</td>
<td>0.1</td>
<td>1.3</td>
<td>119</td>
<td>50</td>
<td>0.68</td>
<td>40</td>
<td>29</td>
<td>318</td>
<td>0.6</td>
<td>0.16</td>
<td>3.1</td>
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</tr>
<tr>
<td>0.01–1.7</td>
<td>0.01–5.7</td>
<td>0.03–7.4</td>
<td>0.06–25</td>
<td>69–424</td>
<td>0.4–144</td>
<td>0.12</td>
<td>40.01–353</td>
<td>0.02–194</td>
<td>107–1093</td>
<td>0.2–35</td>
<td>0.03–1.13</td>
<td>0.5–32</td>
<td></td>
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<tr>
<td>1125</td>
<td>1185</td>
<td>1132</td>
<td>1057</td>
<td>636</td>
<td>1246</td>
<td>687</td>
<td>778</td>
<td>564</td>
<td>645</td>
<td>252</td>
<td>839</td>
<td>747</td>
<td></td>
</tr>
<tr>
<td>FTMA</td>
<td>0.06±0.1</td>
<td>0.3±0.4</td>
<td>0.3±0.6</td>
<td>2.0±2.2</td>
<td>123±64</td>
<td>57±17</td>
<td>4.1±1.6</td>
<td>30±20</td>
<td>22±17</td>
<td>395±249</td>
<td>2.2±4.0</td>
<td>0.15±0.08</td>
<td>6.4±9.6</td>
</tr>
<tr>
<td>0.06</td>
<td>0.1</td>
<td>0.1</td>
<td>1.4</td>
<td>112</td>
<td>57</td>
<td>0.93</td>
<td>29</td>
<td>21</td>
<td>369</td>
<td>0.9</td>
<td>0.14</td>
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<tr>
<td>0.01–1.9</td>
<td>0.02–4.2</td>
<td>0.03–4.9</td>
<td>0.1–16</td>
<td>60–416</td>
<td>29–129</td>
<td>0.10</td>
<td>40.01–353</td>
<td>0.02–194</td>
<td>107–1093</td>
<td>0.2–22</td>
<td>0.02–0.60</td>
<td>0.9–93</td>
<td></td>
</tr>
<tr>
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<td>563</td>
<td>561</td>
<td>519</td>
<td>351</td>
<td>595</td>
<td>375</td>
<td>237</td>
<td>206</td>
<td>357</td>
<td>122</td>
<td>390</td>
<td>401</td>
<td></td>
</tr>
<tr>
<td>TIC</td>
<td>0.7±1.3</td>
<td>2.0±3.1</td>
<td>2.6±4.4</td>
<td>8.5±7.4</td>
<td>186±81</td>
<td>72±19</td>
<td>22.3±40.4</td>
<td>12±4</td>
<td>101±42</td>
<td>651±372</td>
<td>6.7±10</td>
<td>0.23±0.19</td>
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</tr>
<tr>
<td>0.2</td>
<td>0.6</td>
<td>0.8</td>
<td>6</td>
<td>192</td>
<td>74</td>
<td>9.4</td>
<td>13</td>
<td>108</td>
<td>576</td>
<td>2.4</td>
<td>0.15</td>
<td>3.4</td>
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</tr>
<tr>
<td>0.02–7.4</td>
<td>0.03–19</td>
<td>0.05–26</td>
<td>0.3–40</td>
<td>67–418</td>
<td>40–120</td>
<td>0.15–235</td>
<td>8–18</td>
<td>10–168</td>
<td>104–2100</td>
<td>0.2–36</td>
<td>0.10–0.91</td>
<td>1.5–7.3</td>
<td></td>
</tr>
<tr>
<td>69</td>
<td>69</td>
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<td>59</td>
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<td>47</td>
<td>23</td>
<td>67</td>
<td>69</td>
<td></td>
</tr>
</tbody>
</table>

a Mean±1σ; b Median; c Min.-Max.; d Number of data; e The numbers in parenthesis are Mean±1σ calculated with the data set corresponding to SZA≤85°.

The category for the BL air masses was determined based on meteorological parameters and WRF-Chem; those for FTCO and FTMA based on geographical locations of airborne sampling; that for the BB based on hydrogen cyanide (HCN), perchlorethene (C2Cl4), and CO; and that for TIC based on the concentration levels of CO, NOx, and SO2.

the ratios (0.13–0.15) for marine air during PEM-Tropics B (Maloney et al., 2001).

The absolute magnitude of chemical observations was also different with air mass categories. For instance, for the BB, O3 and NOx concentrations were relatively higher with mean levels of 74 and 3.1 ppbv, respectively, higher than those for TIC, but less than those for the BL (79 and 3.7 ppbv, respectively) (Table 2). Mean concentrations of NO and NO2 for the BB were 0.5 and 2.5 ppbv respectively. Except for the BL, the mean NO2 level for the BB was higher than those in other air mass categories by at least a factor of 1.3 and as much as a factor of 13 for the FTCO. For FTCO, the mixing ratios of air pollutants such as O3 and NOx were low in comparison to other air mass categories, with mean values of 51 and 0.3 ppbv. The relatively high NOx levels for the BB are likely to be affected by outflow from MCMA. NOx enrichment was also reported in the forest fire emissions near MCMA during the MIRAGE-Mex campaign on a different airborne platform (Twin Otter), ascribed to the deposition of nitrogen-containing pollutants in the outflow from the MC urban area (Yokelson et al., 2007). The concentrations of peroxy radicals, HO2 and RO2 for the BB were 58 and 159 pptv, respectively, which were factors of 1.3 to 4.8 (except for the BL) and 1.6 to 7.2 higher than those in other air mass categories, respectively. TIC also showed significantly high RO2 concentrations (approximately 100 pptv).

3.2 Photostationary state analysis of NO-NO2 system

The PSS of NO-NO2 systems with different air mass categories is shown in Fig. 3 and Table 3. The former shows the linear regression between NO2/NO and [k1(O3)+k3[HO2]+k4[RO2]]/(NO2) and the latter shows the statistical summary of the PSS parameter (φ). The mean φ values for different air masses are un-weighted averages, in other words each individual φ value is considered with equal weight. The PSS analysis for the air mass types of BB and TIC was excluded due to the limited number of data (n=8 for BB and...
Fig. 3. Photostationary state analysis for the NO-NO$_2$ system during the MIRAGE-Mex field campaign (4 March to 29 March 2006) for different air mass categories (all a, BL b, FTCO c, and FTMA d). The photostationary parameter $\phi$ is the slope of these plots. The upper and lower (dotted) lines represent prediction intervals in 95% confidence level.

Table 3. Statistical summary of the PSS parameter ($\phi$) and linear regression between NO$_2$/NO and $\{k_1[O_3]+k_3[HO_2]+k_4[RO_2]\}/J(\text{NO}_2)$.

<table>
<thead>
<tr>
<th>Air mass category</th>
<th>Slope$^a$</th>
<th>Slope SE$^b$</th>
<th>S$^c$</th>
<th>T-value</th>
<th>P-value</th>
<th>$\Delta \phi / \phi$ (%)</th>
<th>$\phi$ (mean±1σ)</th>
<th>$\phi$ (median)</th>
</tr>
</thead>
<tbody>
<tr>
<td>BL</td>
<td>1.14</td>
<td>0.012</td>
<td>0.75</td>
<td>91.97</td>
<td>0</td>
<td>21</td>
<td>1.19±0.24</td>
<td>1.14</td>
</tr>
<tr>
<td>FTCO</td>
<td>1.30</td>
<td>0.029</td>
<td>2.79</td>
<td>45.50</td>
<td>0</td>
<td>22</td>
<td>1.06±0.37</td>
<td>1.04</td>
</tr>
<tr>
<td>FTMA</td>
<td>0.89</td>
<td>0.018</td>
<td>1.48</td>
<td>50.18</td>
<td>0</td>
<td>21</td>
<td>0.93±0.27</td>
<td>0.92</td>
</tr>
<tr>
<td>All</td>
<td>1.13</td>
<td>0.017</td>
<td>2.42</td>
<td>66.75</td>
<td>0</td>
<td>–</td>
<td>1.08±0.37</td>
<td>1.06</td>
</tr>
</tbody>
</table>

Values were calculated from the data subset with SZA≤85°.

$^a$ Slope of the best-fit line in the Fig. 1. $^b$ Standard error of slope. $^c$ Square root of mean squared error.
Fig. 4. Vertical profile of the PSS parameter (φ) for 3 air mass categories (all a, BL b, FTCO c, and FTMA d).

The PSS parameter (φ) is defined as follows:

\[
φ = \frac{\{k_1[O_3] + k_3[HO_2] + k_4[RO_2]\}}{[NO_2]/[NO]} \tag{2}
\]

In Eq. (2), the reaction rate constants (k₁, k₃, and k₄) are taken from Sander et al. (2002) and k₄ is the coefficient corresponding to the reaction between NO with CH₃O₂. All rate coefficients were adjusted for the temperatures measured during the flights. If Reactions (1–4) represent adequately the NO-NO₂ partitioning, the calculated value of φ (or the slope of Fig. 3) would be expected to be near unity. Note that φ<1 (or φ>1) indicates that the calculated [NO₂] is less (higher) than observed [NO₂] and that concentration levels of observed peroxy radicals (e.g., [HO₂] and [RO₂]) are too low (high) to sustain PSS, implying a large deviation from the PSS. Note that there was no impact of data near sunrise and sunset (SZA > 85°) on the φ calculations and the linear regression due to lack of measurement parameters on a common time base over these periods. The mean PSS parameter (φ) and the slope of the best-fit line in the Fig. 3 were similar, except for FTCO (Table 3). The mean φ values ranged from 0.93 (FTMA) to 1.19 (BL) with the mean value of 1.08 for all data (Fig. 3a). Based on error propagation analysis using measurement uncertainties, the uncertainty (1σ) of the parameter φ is estimated to range from 21 to 22%.

In general, NO₂/NO ratios are near PSS, showing a strong correlation (r²=0.73) with the value of \{k₁[O₃]+k₃[HO₂]+k₄[RO₂]\}/J₂. A slightly higher mean φ value (1.19) for BL was estimated, compared to those (1.06 and 0.93) for FTCO and FTMA. As the key components of NO-NO₂ partitioning, the concentrations of ozone and peroxy radicals for BL were up to a factor of 1.5 and 2.5 higher than those for FTCO and FTMA, respectively. Table 4 shows the contribution of ozone and peroxy radicals to the observed NO₂/NO ratio, where the statistical values were calculated from the data sets with SZA ≤ 85°. The ratios were controlled mainly by ozone levels (e.g., 61–71% contribution).
Fig. 5. Relation between solar zenith angle (SZA) and the PSS parameter (φ) for 3 air mass categories (all a, BL b, FTCO c, and FTMA d).

Table 4. Contribution of $O_3$ and peroxy radicals to the NO$_2$/NO ratios.

<table>
<thead>
<tr>
<th>Air mass category</th>
<th>$k_1[O_3]/J_2^*$</th>
<th>$(k_3[HO_2]+k_4[RO_2])/J_2^*$</th>
</tr>
</thead>
<tbody>
<tr>
<td>BL</td>
<td>3.18±1.40 (3.04)</td>
<td>1.62±0.47 (1.58)</td>
</tr>
<tr>
<td>FTCO</td>
<td>2.33±4.74 (1.17)</td>
<td>1.52±1.36 (1.27)</td>
</tr>
<tr>
<td>FTMA</td>
<td>3.35±2.48 (2.94)</td>
<td>1.37±0.85 (1.28)</td>
</tr>
</tbody>
</table>

Values were calculated from the data set with SZA≤85°.

As noted above, the discrepancy between the observed [NO$_2$]/[NO] ratios (denominator in Eq. 2) and calculated ratios (numerator in Eq. 2) was slightly different with air mass types. For instance, the discrepancy for BL was 12 to 28% higher than that for FTCO (t-test, P-value=1.22×10$^{-6}$) and FTMA (P-value=8.90×10$^{-19}$). In addition, the PSS value for each air mass is statistically different from unity (t-value=10.83 for BL, 3.28 for FTCO, and −3.40 for FTMA). In other words, the PSS values for the BL and FTCO were higher than unity, whereas that for FTMA was lower than unity. Interestingly, the marine atmosphere only showed value less than unity. As shown in Fig. 3, the PSS values in the lower altitude were somewhat higher than unity, whereas those in the higher altitude were lower than unity. One potential of this can be in part vertical iodine levels, which was indirectly supported by vertical CH$_3$I profile showing slightly higher concentrations in the higher altitude (not shown here). The detailed discussion on this is given below.

The discrepancy between the observed and calculated ratios can suggest the possibility of lack of current understanding of the tropospheric chemistry mechanism. The potential role of iodine chemistry in NO$_2$/NO ratio change has been suggested based on field, model, and kinetic studies (Chameides and Davis, 1980; Davis et al., 1996; Knight and Crowley, 2001). Recently, significant impact of iodine chemistry (e.g., reaction of IO with NO producing NO$_2$) on the observed NO$_2$/NO ratio has been reported for the field study at Rishiri island, Japan (Kanaya et al., 2002, 2007). For the source of IO radicals, the photolysis of iodocarbons (CH$_3$I, CH$_2$I$_2$, C$_2$H$_5$I, and CH$_2$ICl), followed by the reaction of O$_3$ and I can produce the IO radical. It has been reported that the source strength for iodocarbons such as CH$_2$I$_2$, C$_2$H$_5$I, and CH$_2$ICl could possibly exceed that for CH$_3$I (Davis et al., 1996 and references therein). In high biological productivity regions, CH$_3$I concentration in the coastal air reached as high as 43 pptv (Oram and Penkett, 1994). When the iodine chemistry scenario is applied to this study, the IO mixing ratio required to reproduce the PSS of NO-NO$_2$ system (i.e., increase of $\phi$ from 0.93 to 1.0) were estimated to be 0.8 pptv for FTMA on average. Similar ranges of IO levels (e.g., 0.3 pptv at Cape Grim and 0.5–7 pptv at Mace Head) have been observed in the marine atmosphere (McFiggans, et al., 2000; Allan et al., 2000; Saiz-Lopez and Plane, 2004). Based on the analysis of air mass back trajectory for FTMA (not shown here), the air mass originated from the Pacific coast of Mexico, which is a region of upwelling of nutrient-rich
seawater (Bulgakov et al., 2005). In the tropical region, there is the strong vertical mixing of CH$_3$I from boundary layer to free troposphere (Davis et al., 1996). Thus, this suggests the potential role of halogen chemistry in NO$_2$/NO ratio shift for the marine atmosphere.

It appears that there is slight trend in the PSS parameter $\phi$ in terms of altitude (above ground level and a.s.l.) (Fig. 4). With increasing altitudes, the proportion of $\phi$ values less than 1 generally increases, especially for FTCO and FTMA. In addition, the relation between $\phi$ and SZA is shown in Fig. 5. There was no clear trend between these two parameters. For the BL, the $\phi$ values were almost constant regardless of the SZA. The $\phi$ values for the FTCO showed large variation both at the ranges from 25 to 30° SZA and from 70 to 85° SZA, whereas they were almost uniform in between. However, $\phi$ values for FTMA varied significantly but not systematically with SZA.

According to a previous study of the PSS analysis of NO-NO$_2$ system (Crawford et al., 1996), the PSS parameter ($\phi$) was reported to range from 0.30 to 0.75, estimated for several airborne sampling campaigns such as PEM-West A, CITE-3, ABLE-3B, CITE-2, and TRACE-A. From these campaigns, most $\phi$ values were close to 0.71, but that for PEM-West A (0.30) was significantly lower in comparison. The cause for the significantly large deviation from the PSS in PEM-West A was suggested to be interference in the NO$_2$ measurement and this possibility stimulated improvements to the measurement technique during PEM-Tropics A (Bradshaw et al., 1999).

3.3 Ozone Production Efficiency (OPE)

Photochemical oxidation of NO$_x$ and its oxidation products were interrelated with O$_3$ formation during the daytime. The OPE, which is the number of O$_3$ molecules produced per molecule of NO$_x$ oxidized (Kleinman et al., 2002 and references therein), was estimated using the linear regression slope between O$_3$ and oxidized NO$_x$ products (NO$_y$=NO$_2$−NO$_x$). Figure 6 shows the correlation between odd oxygen (O$_3$+NO$_2$) and NO$_x$, with OPEs for the 5 air mass categories in this field campaign ranging from 4.5 (TIC) to 8.5 (FTMA), indicating progressively more photochemically aged air masses. In general, higher efficiencies (5.9–8.5) were observed in FTCO and FTMA, which can be supported in part by lower NO$_2$/NO ratios, whereas lower OPEs (4.5–4.6) were found in TIC and BB. To some extent, these lower OPEs in TIC and BB might result from high NO levels (0.5–0.7 ppbv) that are enough to cause enhancements in OH levels and thus shorten NO$_x$ lifetime, and from higher organic peroxy radicals leading to additional NO$_x$ loss through organic nitrate formation. The OPEs in this field campaign were within the OPE ranges for the urban and rural atmosphere (Rickard et al., 2002 and references therein).

4 Summary and conclusions

Previous studies testing the NO-NO$_2$ cyclic system were conducted in urban and remote areas using observations of O$_3$ (R1) and/or peroxy radicals (R3–R4). These earlier studies used the concentrations of model-predicted peroxy radicals to test the PSS for NO-NO$_2$ cyclic system and indicated that the peroxy radicals were not sufficient to explain the observed NO$_2$/NO ratio. In addition, the model-predicted NO$_2$ levels in previous studies were reported to be somewhat lower than observations. In this study, we analyzed the NO-NO$_2$ system in different chemical regimes/air masses based on observations of reactive nitrogen species and peroxy radicals measured during the intensive field campaign of MIRAGE-Mex (4 March to 29 March 2006). For the PSS analysis, the air mass was categorized into 5 groups such as BL, BB, FTCO, FTMA, and TIC.

In general, NO$_2$/NO ratios seem to be near PSS. In addition, the PSS value ($\phi$) for each air mass is statistically different from unity. In other words, the PSS values for the BL and FTCO were higher than unity, whereas that for FTMA was lower than unity. Overall, there was no distinct PSS dependence on altitude and SZA. The mean NO$_x$/NO$_2$ ratios varied from 0.15 (FTMA) to 0.38 (BL), indicating less photochemically aged air masses of polluted plume exiting Mexico City Metropolitan Area. Ozone production efficiencies ranged from 4.5 (TIC) to 8.5 (FTMA). The potential role of halogen chemistry in NO$_2$/NO ratio shift (i.e., lowering the PSS parameter, $\phi$) in the marine atmosphere (FTMA case) can not be excluded according to our simple estimated IO concentration for the conversion of NO to NO$_2$ by IO.

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References


Tie, X., Madronich, S., Li, G. H., Ying, Z., Zhang, R., Garcia, A. R., Lee-Taylor, J., and Liu, Y.: Characterizations of chemical...